Vegetation Development in Abandoned Fallowland in Chapra, Bihar, India

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ABSTRACT

The floristic composition, life form and live shoot biomass estimation of a grassland community developed in abandoned cropland after about nine years at Jai Parkash University Chapra ($25^0 36' \cdot 26^0 15$ 'N lat. and $84^0 24' \cdot 85^0 15$ 'E long.) were studied. The floristic compositoin of the grassland community was composed of 14 species. *Cynodon dactylon* and *Dicanthium* annulatum were the dominant species. Other important species were *Convolvulus arvensis* and *Cyperus rotundus* during the study period. In the study sites cryptophytes, hemicryptophytes and chamaephytes were present however phanerophytes and therophytes were absent. The IVI values ranged from 56.39 to 82.68 for *Cynodon dactylon* and *Dicanthium annulatum*, respectively. The shoot biomass values ranged from 26.83 to 41.32 gm⁻² in two study sites. The species richness and equitability values ranged from 1.64 to 2.73, and 2.22 to 2.73, respectively at two study sites.

Keywords: Abandoned cropland, Species composition, Importance value index, Species richness, Equatability and Live shoot biomass

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INTRODUCTION

Ecologically a grassland may be defined to be the land on which graminoid (poaceae) species are dominant. Grassland as a whole encomposses the potential natural vegetation of 25% of the earth's land surface (Shantz 1954), and accounts for about 16% (1.89×10^{10} tha⁻¹) of the net primary productivity of plant community (Whittaker and Likens, 1973). According to Singh (1987), grasslands in India occupy 39.81 per cent (12121 thousand ha) area of the total Indian sub-Chanda, *et al.*, 2016: Vol 4(4) 156

continent.Grasslands in India have developed either due to deforestation or abandoned cultivation (Misra 1983, Blasco et al. 1983, Singh et al. 1985). Whyte (1968) recogonized that grasslands occur on almost all soil types and their distribution is predominently governed by climatic factors. The tropical grasslands in India are seral in nature due to reccuring biotic operation such as grazing, fire and scraping (Neeraj et al. 2004). Mankind has depended upon grasslands for his existance even since our remote ancestors ventured onto the savannas and began a new mode of existance. Grasses seem to have emerged during the late Cretaceous and by early Miocene and the grasslands probably assumed a prominent place in the earth's vegetation . Most works on grassland ecosystems in India have been reported by Pandeya et al. (1977), Singh and Yadava (1974), Sinha et al. (1991), Ramakrishnan and Ram (1988), Mishra and Mishra (1984) etc.

Since 2006 cultivation of crops has been banned in the J. P. University Chapra, campus and the campus is now a fallowland. Only within nine years the vegetation has developed which is a grassland type. According to Bignal (1996) and Kleijn et al. (2006) some agricultural and agroforestry systems that shape cultural landscape have been recogonized which need conservation relevance includiing biodiversity, habitat and aesthetic values. The variable ecosystems would be lost if agricultural use is abandoned. There are negative as well as positive effects of abandonment of agricultural land.Following are the five main problems related to negative effects of abandonment of cultivation as discussed by Jose et al. (2007) : a) reduction of landscape heterogeneity and promotion of vegetation homogenisation associate and incresead fire frequency, b) soil erosion and desertification, c) reduction of water stocks, d) biodiversity loss and reduced population of adopted species, and e) loss of cultural and aesthetic values.

When abandonment is simultaneous for large extension of farmlands, vegetation homogenization increases (Lasanta-Martineza et al. 2005) and landscape heterogeneity reduces (Hochtl et al.2005). Reduced landscape heterogeneity increases the disturbances (Lloret et al. 2001). Soil erosion is a problem linked to land abandonment in some parts of the world, but not everywhere . The abandonment of pastoral practices had some negative consequences, such as the penetration of invasive species and an increase of unaltered litter on the ground, which leads to the worsening of the pastoral features (Susan et al. 2004). The abandonment of farmland and its detrimental effects on landscape heterogeneity result in the loss of plants (Luoto et al. 2003 and Laiolo et al. 2004) and invertibrates (Inoue et al. 2005 and Daube et al. 2006). The abandonment of agricultural land may also have a variety of positive consequences such as : a) revegetation and forest plantations ,b) water retension and soil recovery, c) nutrient cycling and d) increase in biodiversity (Jose et al. 2007).

Due to colonisation of abandoned land plants and animals can disperse from surrounding habitats and subsequently establish, survive and grow. It integrates natural condition with plant cover. It results increase in the density and distribution of biomass. Rapid ecosystem recovery following cropland abandonment and rural - urban migration has been documented in forested and non forested ecosystems in many regions of the world (Lepers et al. 2005, Aide et al. 2005, Grau et al. 2007 and Vallejo et al. 2006). There are a lot of benefits obtained from revegetation of abandoned agricultural land. These include hydrological regulation, soil recovery and erosion mitigation, increased fertility and higher water quality. Due to sucessional vegetation development higher evapotranspiration and infiltration rates and therefore runoff is reduced and an increase in water holding capicity takes place (Sponge effect) (Bruijnzeel et al. 2004). Soil properties improve when afforestation practices after land abandonment are successful. Landscape heterogeneity may decrease or increase by the abandonment agricultural land affecting the habitat diversity. Abandonment of agricultural lands promotes population decline of species adapted to open space, however it favours the population and diversity of species characteristic of woody vegetation habitats (Laiolo et al. 2004, Muniz et al. 2006, Dunn et al. 2004 and Otto et al. 2006) and of soil fauna (Kardol et al. 2005). Abandonment leads to a high increase in biodiversity compared with farmland (Dunn et al. 2004 and Kappele et al. 1995). The present study was aimed to know the species composition and live shoot biomass level of grassland vegetation developed after nine years of abandonment of cultivation in the Jai Parkash University campus, Chapra, Bihar, India. Data on such studies are not available in the Indo –Gangetic plains of Bihar.

MATERIALS AND METHODS

This study was conducted at two openly grazed vegetation sites (Site I and Site II). Site I was moderately grazed and site II was intensively grazed. The sites were situated between $(25^0 36^2)$ $26^{\circ}15$ 'N lat. and $84^{\circ}24$ '- $85^{\circ}15$ 'E long.) The maximum tempreture values ranged from $15.4^{\circ}to44.5^{\circ}$. Harvest method of Odum (1960) was employed for the estimation of plant biomass. Ten quadrats of 50cmx50cm sizes were randomly harvested 1cm above the ground surface during the last week of August 2015. The samples were packed in polyethylene bags separately and brought to the laboratory for identification (Muller Dombois and Ellenberg 1974). Identification of all the species were made in consultation with various regional and national flora books i,e The Botany of Bihar and Orrisa (Haines, 1921-25); Supplement to the Botany of Bihar and Orrisa (Mooney 1950). The phytosociological observations were made. The species present in quadrats were noted and their numbers were counted individually/tiller of each species. The vegetation data were quantatively Chanda, et al., 2016: Vol 4(4)

analysed for frequency, density, abundance and dominance following Curtis and McIntosh (1950). Relative ferquency, relative density and relative dominance were determined following Phillps (1959), for dominance shoot biomass data were used. The Importance Value Index (IVI) was the sum of relative frequency, relative density and relative dominance. Following formulae were used (Misra 1968):



ns =Biomass value of least important species

Species richness = S-1/In N

S=Total number of species present

N = Total biomass of all species

RESULTS

The present study deals with the phytosociological attributes of an openly grazed vegetation developed in two sites (I and II) after nine years of abandonment of cultivation in the University campus of Jai Prakash University, Chapra. The maximum number of species were recorded 14 at site I whereas this value was 9 at site II. Altogether following 14 species were recorded in both sites:

Aristida cyanantha , Boerhaavia diffusa , Convolvulus arvensis, Convolvulus pluricaulis, Cyperus alulatus, Cynodon dactylon, Cyperus dubius, Cyperus rotundus, Desmodium triflorum, Dicanthium annulatum, Evolvulus alsinoides, Evolvulus nummularius, Parthenium hysterophorus and Sida cordifolia.

Frequency : The values for frequency ranged from 20 to 90 (%) at site I and from 10-100(%) at site II (Tables 1 and 2). In site I maximum value (90%) was recorded for *Convolvulus rvensiss*, *Cynodon dactylon*, *Desmodium annulatum* and minimum (20%) *for Convolvulus pluricaulis*, *Cyperus rotundus*, *Desmodium triflorum*, *Evolvulus alsinoides*, *and Sida cordifolia*. In site II maximum value (100%) was recorded for *Convolvulus arvensis*, *Cyperus dubius* and *Dicanthium annulatum* and minimum (10%) for *Evolvulus alsinoides*.

Density : The values for density ranged from 2.8to 351.2 m^{-2} in site I and from 2.4-606.4 m⁻² in site II (Tables 1 and 2). In site I maximum value (351.2 m^{-2}) recorded for *Cynodon dactylon* and minimum (2.8m^{-2}) for *Convolvulus pluricaulis* and *Desmodium triflorum*. In site II maximum value (606.4m^{-2}) was recorded for *Dicanthium annulatum* and minimum(2.4 m^{-2}) for *Parthenium hysterophorus*.

Abundance : The values for abundance ranged from 11.3to 820 in site I and from 24-606.4 in site II (Tables 1 and Table 2). In site I maximum value (820) was recorded for *Cyperus rotundus* and

minimum (11.3) for Parthenium hysterophorus. In site II maximum value (606.4) was recorded for Dicanthium annulatum and minimum (24) for Parthenium hysterophorus.

Dominance(D₀): The live shoot biomass values ranged from 0.1to 26.83 gm⁻² in site I and from 0.71-41.32gm⁻² in site II for individual species (Tables 1 and 2). In site I maximum value (26.83gm⁻²) was recorded for *Cynodon dactylon* and minimum (0.1gm⁻²) for *Desmodium triflorum*. In site II maximum value (41.32 gm⁻²) was recorded for *Dicanthium annulatum* and minimum (0.71 gm⁻²) for *Parthenium hysterophorus*.

Relative frequency: The values for relative frequency ranged from 2.66-12% in site I and from 1.66-16.66% Π Ι in site .In site maximum value (12%)was recorded for Convolvulus arvensis, Cynodon dactylon and Dicanthium annulatum and minimum (2.66%) for Convolvulus pluricaulis, Cyperus rotundus and Desmodium triflorum. In site II maximum value (16.66%) was recorded for Convolvulus arvensis, Cyperus dubius and Dicanthium annulatum and minimum (1.66%) for Parthenium hysterophorus.

Relative density : The values for relative density ranged from 0.17 to 21.53 % in site I and from 0.13-34.17 % in site II (Tables 1 and 2). In site I maximum value (21.53%) was recoded for *Cynodon dactylon* and minimum (0.17%) for *Convolvulus pluricaulis* and *Desmodium triflorum*. In site II maximum value (34.17%) was recorded for *Dicanthium annulatum* and minimum (0.13%) for *Parthenium hysterophorus*.

Relative dominance(RD_0): The values for relative dominance ranged from 0.08to 22.86 % in site I and from 0.55-31.85 % in site II (Tables1 and 2). In site I maximum value (22.86%) was recorded for *Cynodon dactylon* and minimum (0.08%) for *Desmodium triflorum*. In site II maximum value (31.85%) was recorded for *Dicanthium annulatum* and minimum (0.55%) for *Parthenium hysterophorus*.

Important Value Index(IVI): IVI values ranged from 2.83 to 56.39 % in site I and from 2.34-82.68 % in site II(Tables 1 and 2). In site I maximum value (56.39) was recorded for *Cynodon dactylon* and minimum (2.83) for *Convolvulus pluricaulis*. In site II maximum value (82.68) was recorded for *Dicanthium annulatum* and minimum (2.34) for *Parthenium hysterophorus*.

 Table : 1 The phytosociological characteristics of vegetation in fallowland developed after nine years of abandonment of cropping at site I

Species	F(%)	D(m ⁻²)	А	D ₀ (gm	RF(%)	RD(%)	RD ₀ (%)	IVI
				²)				
<u>Aristida</u> <u>cyanantha</u>	80	145.6	182	8.16	10.66	8.92	6.96	26.54
<u>Boerhaavia</u> <u>diffusa</u>	60	51.2	85.3	5.7	8	3.14	4.86	16
<u>Convolvulus</u>	90	288.4	320.4	16.80	12	17.69	14.32	44.01
<u>arvensis</u>								
<u>Convolvulus</u>	20	2.8	14	0	2.66	0.17	0	2.83
<u>pluricaulis</u>								
<u>Cyperus</u> alulatus	40	84	210	6	5.4	5.15	5.11	15.66
<u>Cynodon</u> dactlon	90	351.2	390.2	26.83	12	21.53	22.86	56.39
<u>Cyperus</u> <u>dubius</u>	80	216.4	270.5	11.58	10.66	13.26	9.87	33.79
<u>Cyperus</u> rotundus	20	164	820	9.57	2.66	10.05	8.15	20.86
<u>Desmodium</u>	20	2.8	14	0.1	2.66	0.17	0.05	2.88
<u>triflorum</u>								
<u>Dicanthium</u>	90	262.8	292.0	21.83	12	16.11	18.40	46.51
<u>annulatum</u>								
<u>Evolvulus</u>	20	32	160	3.7	2.66	1.96	3.15	7.77
<u>alsinoides</u>								
<u>Evolvulus</u>	60	16.4	27.3	1.82	8	1.01	1.55	10.55
<u>nummularus</u>								
<u>Parthenium</u>	60	6.8	11.3	4.98	8	0.41	4.24	12.65
<u>hysterophorus</u>								
<u>Sida</u> cordifolio	20	6.4	32	0.46	2.66	0.39	0.39	3.44

 $F = Frequency, D = Density, A = Abundance, D_0 = live shoot biomass, RF = Relative frequency, RD = Relative density, RD_0 = Relative dominance, IVI= Importance Value Index$

 Table : 2 Phytosociological characteristics of vegetation in fallowland developed after nine years of abandonment of cropping at site II

Species	F(%)	D(m ⁻²)	A	D₀(gm ⁻ 2)	RF(%)	RD(%)	RD ₀ (%)	IVI
<u>Aristida</u> <u>cyanantha</u>	90	184.8	205.3	9.70	15	10.41	7.48	32.89
<u>Boerhavia</u> diffusa	-	-	-	-	-	-	-	-
<u>Convolvulus</u> <u>arvensis</u>	100	206.8	206.8	25.78	16.66	11.65	19.8	48.11
<u>Convolvulus</u> <u>pluricaulis</u>	-	-	-	-	-	-	-	-
<u>Cyperus</u> <u>alulatus</u>	-	-	-	-	-	-	-	-
<u>Cynodon</u> <u>dactylon</u>	90	177.2	196.8	11.53	15	9.98	8.88	33.86
<u>Cyperus</u> dubius	100	479.6	479.6	28.09	16.66	27.02	21.65	65.33
<u>Cyperus</u> rotundus	-	-	-	-	-	-	-	-
<u>Desmodium</u> <u>triflorum</u>	50	46.8	93.6	6.45	8.33	2.63	4.97	15.93
<u>Dicanthium</u> <u>annulatum</u>	100	606.2	606.4	41.32	16.66	34.17	31.85	82.68
<u>Evolvulus</u> <u>alsinoides</u>	20	5.6	28	0.75	3.33	0.31	0.58	4.22
<u>Evolvulus</u> <u>nummularis</u>	40	64.8	162	5.38	6.66	3.65	4.14	14.45
<u>Parthenium</u> <u>hysterophorus</u>	10	2.4	24	0.71	1.66	0.13	0.55	2.34
Sida cordifolia	-	-	-	-	-	-	-	-

F = Frequency, D = Density, A = Abundance, $D_0 =$ live shoot biomass, RF = Relative frequency,

RD= Relative density, RD_0 = Relative dominance, IVI= Importance Value Index.

<u>Dicanthium</u> annulatum had maximum importance at both the sites followed by <u>Cynodon</u> <u>dactylon</u> at sites I and <u>Cyperus</u> <u>rotundus</u> at sites II. The species richness values were 2.73 and 1.64 at sites I and sites II, respectively. The equitability value ranged from 2.22 to 2.51 (Table 3).

Table: 3 Spec	ties composition	of grassland	community a	t Site I	and Site II
	1				

Variables	Site I	Site II
Total number of species	14	9
Species common to both sites.	9	
Species exclusive to either sites.	5	00
Equitability	2.51	2.22
Species richness	2.73	1.64

DISCUSSION

The present investigation deals with the floristic composition of the vegetation in rainy season (Site I and Site II) in Jai Prakash University campus , Chapra. The herbaceous grass-domminated community of study sites contain high percentage of hemicryptophytes, cryptophytes and chaemophytes (Table 4) that appears as a result of grazing and browsing in the area .Generally grasslands are rich in hemicryptophytes (Braun Blanquet, 1932). Maximum percentage of hemicryptophytes and absence of phanerophytes and therophytic species attributed to be due to influence of periodicity, soil characterestics, climatic condition as well as the biotic interferences (Table 5). Similar findings were also reported in grasslands of India (Singh and Ambasht 1975,Abdar, 2013, Bark and Misra, 1997). In the present study at two sites relatively higher number of species at site I can be attributed to moderate grazing. The much higher abundance and Chanda, *et al.*, 2016: Vol 4(4)

dominance of species like *Aristida cyanantha*, *Convolvulus abinoids*, *Cyperus rotundus* and *Dicanthium annulatum* at site II may be related to their characteristic growth pattern showing profuse sprouting and flowering even at highly disturbed sites.

Species	Life form occurred
<u>Aristida</u> <u>cyanantha</u>	Не
<u>Boerhaavia</u> <u>diffusa</u>	Не
Convolvulus arvensis	Не
<u>Convolvulus</u> <u>pluricaulis</u>	Не
<u>Cyperus</u> <u>alulatus</u>	Не
<u>Cynodon</u> <u>dactylon</u>	Не
<u>Cyperus</u> <u>dubuis</u>	Не
<u>Cyperus</u> <u>rotundus</u>	Cry
Desmodium triflorum	Не
<u>Dicanthium</u> annulatum	Не
Evolvulus alsinoides	Не
<u>Evolvulus nummularus</u>	Cha
Parthenium hysterphorus	Не
<u>Sida</u> cordifolia	Cry

Table: 4 Life forms of different	species in	study sites
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Region	Pha(%)	Cha(%)	Hem(%)	Cry(%)	The(%)
Raunkiaer's normal spectrum (Raunkier, 1934)	16	9.0	26	6	43
Varanasi (Rao, 1968)	40.0	6.0	1.0	10.0	43.0
Varanasi (Singh and Ambasht, 1975)	-	4.2	19.2	6.3	70.2
Berhampur (Misra and Misra, 1979)	5.7	25.7	14.3	5.7	48.6
Western Orrisa (Naik, 1985)	3.0	21.20	18.20	6.0	51.50
South Orrisa (Patnaik, 1993)	3.58	17.86	25.00	10.71	42.86
Present study	-	7.14	78.57	14.28	-

Pha =Phanerophytes, Cha =Chamaephytes, Hem= Hemicryptophytes, Cry= Cryptophytes, The = Therophytes

The frequency values ranged from 20 to 90 % at sites I to 20 to 100% at sites II.In site I maximum value (90%%) was recorded for Cynodon dactylon and minimum value (20%) for Convolvulus pluricaulis. In site II maximum value (100%) was recorded for Dicanthium annulatum and minimum value(20%) for Evolvulus alsinoides. The density values ranged from 2.8to 351.2 m⁻² at site I and 2.4to 606.2 m⁻² at site II.However in Kurushetra density values ranged from 471to 2143 m⁻² (Singh and Yadava, 1974), in Varanasi values ranged from 1963 to 11055 m^{-2} (Singh 1967), in Ujjain values ranged from 104 to 862 m^{-2} (Misra 1973), and in Ratlam values ranged from 21to 4700 m⁻² (Billore 1973). The IVI values ranged from 2.83 to 56.39 at site I and from 2.34 to 82.68 at site II. On the basis of IVI values in the present study at site I C. dactylon -D. annulatum community and at site II D. annulatum – C. dubius community have developed after nine years of abandonment of cropland. In a study of grassland vegetation at Kurushetra, Singh and Yadava (1974) noted that some species exhibit maximum density during rainy season (e.g. Bothriochloa pertusa and Sorghum halepens). The live shoot biomass values ranged from 0.1 to 26.83 gm⁻² and site I to 0.7to 41.32 at site II. Pandeya (1974) has reported live Chanda, et al., 2016: Vol 4(4) 166

shoot biomass value for *Cenchrus ciliaris* dominated grassland from 0 to 228 gm⁻², at Rajkot, India. Kumar and Joshi (1972) have reported live shoot biomass values for mixed grass dominated grassland from 35 to 76 gm⁻² at Pilani,India ; 105 to 1974 gm⁻² at Kurshetra (Singh and Yadava 1974) ,24 to 457 gm⁻² at Ujjain (Misra 1973), 1 to 363 gm⁻² at Ratlam(Billore 1973), 14 to 572gm⁻² at Sagar for *Heteropogon* dominated grassland (Jain1971).

C. dactylon : It is also known as Doob grass. The blades are a grey – green and are shorts, usually 2-15 cm long with rough edges. It is fast growing and tough and useful for sports fields, as when damaged it will recover quickly. Doob grass has been cultivated on saline soils. It is also highly aggressive, crowding out most other grasses and invading other habitats, and has become a hard to-eradicate weed in some areas.

D. annulatum: The species is quite variable. The stems spread outwards, then grow erect at the ends. Each node on the stem is enriched with a ring of hairs. It can be used in fields for grazing, livestock, and cut for hay and silage. It forms a tuft that can stand up to grazing pressure.

C. dubius : This species is distributed in old world tropics . It is a perennial herb growing in seasonally flooded areas and in pockets of soil in rocks. It is not confined to wetlands and in sometimes found as a weed in fields and near the sea. It is also seen in open shady places.

CONCLUSION

In the moderate grazing site I had more number of species than the intensively grazed site II. Intensive grazing decreased the total number of species. The life form studies indicated a hemi-cry chaemephytic flora. The live shoot biomass values in growing season is less than the other studies conducted in India. *P. hysterophorus* is invading the study sites aggressively affecting the plant diversity in the study sites . Thus to accelerate the natural course of succession intensive grazing should be checked and eradication of *Parthenium*_ is needed.

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REFERENCES

Aide, T.M. and Grau, H. R. (2005) : Globalisation, migration, and Latin American ecosystem. *Science*, 305 : 1915-6.

Abdar. (2013): Physico- chemical characteristics and phytoplankton of Morna Lake, Shirala (M. S) India, *Biolife*, 1 (2): 1-7.

Barik, K. L. and Misra, B .N (1997): Seasonal variation in the standing crop biomass of an upland grassland community. *Mendle*, 14(1 & 2): 9-10.

Bignal, E.M. and McCracken, D. I. (1996) : Low – intensity farming systems in the conservation of the countryside. *Journal of Applied Ecology*, 33 : 413-24.

Billore, S. K. (1973): Net primary production and energetics of a grassland ecosystem at Ratlam, India. Ph.D. Thesis, Vikram University, Ujjain, India.

Braun – Blanquet, J. (1932), Plantsociology: The study of plant communities (translated by Fuller, G.D. and Conrad, H. S.) .Mc Graw Hill, N. Y.

Bruijnzeel, L. A (2004): Hydrological functions of tropical forests : not seeing the soil for the trees, *Agriculture, Ecosystems and Environment*; 104: 185-228.

Curtis, J. T. and McIntosh, R. P.(1951): An upland forest continuum in the praire forest border region of Wisconsin. *Ecology*, 32: 476- 496.

Dauber, J., Bengtsson, J. and Lenoir, L. (2006) : Evaluating effects of habitat loss and land – use continuity on ant species richness in semi natural grassland remnants. *Conservation Biology* ; 20: 1150-60.

Dunn, R. R.(2004): Recovery of faunal communities during tropical forest regeneration . *Conservation Biology*; 18: 302-9.

Grau, R. H, Perez Ceballos, M., Martinuzzi, S, Encarnocion, X. and Aide, T. M. (2007): Cambios socioeconomicos y regeneration delbosque en la Republica Dominicana. In: Gonzales-Espinosa M, Rey- Benayas, J.M., Ramirez-Marcial, N., editors Restauracion de bosques en American Latina. Ediciones Mundi-Prensa y Fundacion Internacional para la Restauraion de Ecosistemas. Mexico D.F., Mexico. In press.

H. L. Shantz (1954): The place of grasslands in the earth's cover. *Ecology*; 35 : 143-145.

Haines, H.H. (1961): The Botany of Bihar and Orrisa.Vol. I to VI, Landon . *Botanical Survey of India*. Calcutta (Rep. Edn.).

Hochtl, F., Lehringer, S. and Konold, W. 'Wilderness' (2005) : What it means when it becames a reality – a case study from the southwestern Alps. Landscape and Urban Planning ; 70: 85-95.

Inoue,T.(2005): Cause of butterfly decline in Japan. *Japanese Journal of Entomology* (New Series); 8: 43-64.

Jain, H.K.(1971): Production studies in some grasslands of Sagar. Ph. D. Thesis. Saugar University, Sagar, India.

Jose, M., Rey Benayas^{*}, Ana Martins, Jose, M., Nicolau and Jennifer, J., Schulz; (2007): Abandonment of agricultural land : an overview of drivers and consequences ; *Perpectives in Agriculture, Vaterinary Science, Nutrition and Natural Resources*, 2 (057) : 1-13.

Kappelle, M., Kennis, P.A.F. and deVries, R.A.J.(1995): Changes in diversity along a successional gradient in a Costa Rica upper montane quercus forest. *Biological Conservation*; 126: 317-327.

Kardol, P., Bezemer, T.M., and van der wal, A., van der Putten, W. H.(2005): Successional trajectories of soil nematode and plant communities in a chronosequence of ex-arable lands. *Biological Conservation*; 126: 317-327.

Kleijn, D., Baquero, R.A., Clough, Y., Diaqz, M., De Esteban, J. and Fernandez, F., et al. (2006) : Mixed biodiversity benefits of agrienvironment schemes in five European countries . *Ecology Letters*; 9 : 243-54.

Kumar, A., and Joshi, M. C.(1972) : The effect of grazing on the structure and productivity of the vegetation near Pilani Rajasthan, India. *Journal of Ecology*. 60: 665-674.

Lasanta- Martineza, T., Vicente- Serrano, S. M. and Cuadrat-Parts, J. M. (2005): Mountain Mediterraanean landscape evolution caused by the abandonment of traditional primary activities : a study of the Spanish Central Pyrenees. *Applied Geography*; 25 : 47-65.

Laiolo, P., Dondero, F., Cilliento, E. and Rolando, A.(2004) : Consequences of pastoral abandonment for the structure and diversity of the alpine avifauna. *Journal of Applied Ecology*; 41: 294-304.

Lepers, E., Lambin, E. F., Janetos, A.C., DeFries., Achard, F. and Ramankutty, N., et al. (2005) : A synthesis of rapid land – cover change information for the 1981-2000 period. *BioScience*; 55: 19-26.

Lloret, F. and Mari, G. (2001) : A comparision of the medieval and the current fire regimes in managed pine forests of Catalonia (NE Spain). *Forest Ecology and Management*; 141 : 155-63.

Luoto, M., Pykala, J. and Kuussaari, M.(2003) : Decline of landscape –scale habitat and species divertsity after the end of cattle grazing *Journal for Nature Conservation*; 11 :171-8.

Misra, R. (1968): Ecology work book . Oxford and IBH Publication Co; New Delhi.

Misra, C. M. (1973): Primary productivity of a grassland ecosystems at Ujjain, Ph. D. Thesis. Vikram University, Ujjain.

Misra, M. K. and Misra, B.N. ,(1984) : Biomass and primary production in an Indian grassland. *Tropical Ecology*; 22(1) : 88-98.

Mooney, H.F.(1950) : Supplement to the Botany of Bihar and Orrisa. Catholic Press., Ranchi.

Mueller – Dombois, D. and Ellenberg, H. (1974) : Aims and methods of vegetation ecology, John. Willey and Sons ; New York, 547p.

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Muniz, M. A., William-Linera and G. and Rey-Benayas, J. M. (2006) : Distance effect from cloud forest fragments on plant community structure in abandonment pastures in Veracruz, Mexico . *Journal of Tropical Ecology*; 22 : 431-40.

Neeraj, S. R., Gupta, V., Malik, B., Kaur and Neelam. (2004): Plant diversity, carbon dynamics and soil biological activity in tropical successional grassland systems at Kurushetra. *International Journal of Ecology and Environmental Science*; 30: 285-298.

Otto, R., Krusi, B. O. and Burga and C. A., Fernandez-Palacois, J. M. (2006) : Old field succession along a precipitation gradient in the semi- arid coastal region of Tenerife. *Journal of Arid Environments*; 65: 156-78.

Odum, E. P.,(1960): Organic production and turnover in old field succession. Ecology, 41: 39-49.

Pandeya, S. C. ,Mankad, M. R. and Jain , H. K. ,(1974) : Potentialities of net primary production of arid and semi – arid grazing lands of India . Proceedings of the seventh *International grassland Congress.*,136-170.

Pandeya, S. C., Sharma, S. C., Jain, H. K., Pathak, S.J., Paliwal, K. C. and Bhanot, V. M. (1977): The environment of <u>*Cenchrus*</u> grazing lands in Western India. Final report of the research project on Genecology and Autecology of Anjan grass (<u>*Cenchrus*</u> <u>*ciliaris*</u>) complex in Western India, Department of *BioScience*, Saurashtra University, Rajkot, India, 451.

Phillips, E. A. (1959): Methods of vegetation study. A. Holtdryden Book Henergy Holt and Congress Inc., 107.

Ramakrishnan, P. S. and Ram, S. C. (1988) : Vegetation biomass and productivity of seral grassland of Cherrapunji, North East, *Vegetatio*; 74: 47-53.

Singh, J. S. (1967): Seasonal variation in composition , plant biomass and net community production in the grassland of Varanasi, Ph. D. Thesis, Banaras Hindu University, Varanas, India.

Singh, J. S. and Yadava, P. S. (1974): Seasonal variation and composition plant biomass and net primary productivity of a tropical grassland at Kurushetra, India. Ecological Monograph ; 44: 351-376.

Singh, U. N. and Ambasht, R. S. (1975): Biotic stress and variability in structure and organic (net primary) production of grassland community of Varanasi, India. *Tropical Ecology*.; 16: 86-95

Singh, P. (1987): Rangeland reconstruction and management for optimising biomass production . *National Rangeland Symposium*, Jhansi, pp 9-12.

Sinha, A. Rana, R. S. and Gupta, S. R. (1991): Growth patterns, net primary production and energy transfers in two grassland communities of sodic soils. *Tropical Ecology*; 32: 105-116.

Susan, F. and Ziliotto, U. (2004) : Effect of arboreal and shrubby components on some features of pastures in a dolomitic area Land use systems in grassland dominated regions. Proceedings of the 20th General Meeting of the European Grassland Federation, 21-24 June 2004, Luzern, Switzerland.

Vallejo, R., Aronson, J., Pausas, J. G. and Cortina, J. (2006) : Restoration of Mediterranean woodlands. In: van Andel , J. , Aronson, J. , editors.Restoration Ecology. : The New Frontier. *Blackwell Science*, Oxford, UK.

Whittaker, R. H. and Likens, G. E. (1975) : The biosphere and man (pp. 305-328) : In primary productivity of the Biosphere (Eds.) Lieth, R. H. and Whittaker R. H. (Eds). Springer-Verlag, New York; 339 pp.

Whyte, R. O., (1968) : Grassland of the Monsoon. (Faber and Faber: London).